

Molecular effects of neonicotinoids in honey bees (*Apis mellifera*)

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1 **Molecular effects of neonicotinoids in honey bees (*Apis mellifera*)**

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31 **Abstract**

32

33 Neonicotinoids are implicated in the decline of honey bee populations. As agonists of the nico-
34 tinic acetylcholine receptors, they disturb acetylcholine receptor signalling leading to neurotox-
35 icity. Several behavioural studies showed the link between neonicotinoid exposure and ad-
36 verse effects on foraging activity and reproduction. However, molecular effects underlying the-
37 se effects are poorly known. Here we elucidated molecular effects at environmental realistic
38 levels of three neonicotinoids and nicotine and compared laboratory studies to field exposures
39 with acetamiprid. We assessed transcriptional alterations of eight selected genes in caged
40 honey bees exposed to different concentrations of neonicotinoids acetamiprid, clothianidin,
41 imidacloporid, thiamethoxam, as well as nicotine. We determined transcripts of several targets,
42 including *nicotinic acetylcholine receptor α 1* and *α 2 subunit*, the multifunctional gene *vitello-*
43 *genin*, immune system genes *apidaecin* and *defensin-1*, stress-related gene *catalase* and two
44 genes linked to memory formation, *pka* and *creb*. *Vitellogenin* showed a strong increase upon
45 neonicotinoid exposures in the laboratory and field, while *creb* and *pka* transcripts were down-
46 regulated. The induction of *vitellogenin* suggests adverse effects on foraging activity, whereas
47 *creb* and *pka* down-regulation may be implicated in decreased long-term memory formation.
48 Transcriptional alterations occurred at environmental concentrations and provide an explana-
49 tion for the molecular basis of observed adverse effects of neonicotinoids to bees.

50

51

52 Introduction

53 Recent reports on global pollinator declines^{1,2} are alarming, especially with respect to the in-
54 creasing demands for pollination services.³ Honey bees are the most economically valuable
55 pollinators⁴. However, the number of managed honey bees decreased by one fourth in Europe
56 between 1985 and 2005⁵, and by more than one half in North America between 1947 and
57 2005.⁵ This decline represents a major challenge for beekeepers and scientists. The reasons
58 are poorly understood. Several studies highlighted the impact of pathogens,⁷ pesticides⁸ and
59 also the lack of wild flowers. Declines similarly occur for wild bees and bumblebees. Multiple
60 chemical residues have been detected inside honey bee hives.⁹ However, no individual factor
61 such as lack of flowers, pesticides or pathogens seems to act as principal driver of colony col-
62 lapse disorder or other honey bee losses. Thus, the global decline of bee populations can be
63 considered as a multifactorial phenomenon driven by a combination of parasites, pesticides
64 and shortage of (wild) flowers.¹⁰ The decline of bee populations has significant negative impli-
65 cations for plant pollination, including many domesticated crops, and therefore, limitation of
66 crop yields.¹⁰

67 Pesticides are currently a debated cause of bee declines. Over the years, the classes
68 of pesticide used in agriculture and their application methods have shifted substantially. Car-
69 bamates, pyrethroids and organochlorides have been less used for the favour of new classes
70 of systemic insecticides, in particular neonicotinoids. Three neonicotinoids, thiamethoxam,
71 imidacloprid and clothianidine, and two organophosphates (phosmet and chlorpyrifos), are
72 thought to pose the highest risk to honeybees on a global scale.¹¹

73 Neonicotinoids are commonly applied as seed-coatings to limit contact with insects and
74 prevent losses. They are neurotoxins that target the central nervous system. Mainly acting as
75 specific agonists by binding to acetylcholine receptors, neonicotinoids are highly effective in
76 disrupting central nervous system function by overstimulation and paralysis, and are thus effi-
77 ciently used for controlling insect pests.¹² In addition to very high toxicity, systemic compounds
78 like neonicotinoids can be particularly problematic for pollinating insects through exposure to
79 residues in nectar and pollen of treated crops.⁸ Recent studies have demonstrated that the
80 hive products of honey bee colonies located in agricultural environments across Europe and
81 North America are contaminated by various pesticides, including neonicotinoids.¹³ Further-
82 more, neonicotinoids are highly persistent in soil and soil water.¹³

83 Neonicotinoids show high acute toxicity to honey bees. Particularly the nitro-substituted
84 compounds, clothianidin, imidacloprid and thiamethoxam (metabolized to clothianidin in plants
85 and insects),¹⁴ show very high toxicity with LD₅₀ values in the range of a few ng/bee. The tox-
86 icity of cyano-substituted neonicotinoids, which include acetamiprid and thiacloprid, is lower
87 with LD₅₀ values in the range of µg/bee.^{15, 16} There exists a considerable variability in inter-
88 individual sensitivity and between colonies.¹⁷

89 Sublethal concentrations of neonicotinoids negatively affected locomotion, behaviour,
90 learning and memory of bees. Imidacloprid, thiamethoxam and clothianidin induced flight mus-
91 cle paralysis,¹⁸ negatively affected learning,¹⁹ olfactory performance, and foraging behaviour
92 ²⁰. Similarly, thiamethoxam ^{21, 22} and acetamiprid ²² decreased memory and olfactory learning
93 capacity. Importantly, neonicotinoids negatively affected orientation and foraging of worker
94 bees. ^{15, 23, 24} These effects were also observed in semi-field conditions. Adverse effects of
95 neonicotinoids on wild bees were shown under field conditions. Rape seed coated with clothi-
96 anidin reduced wild bee density, solitary bee nesting, and bumblebee colony growth and re-
97 production.²⁵ Field realistic doses of imidacloprid alters foraging activity and decreases avoid-
98 ance of predators of honey bees.²⁶ Bumble bee colonies exposed to clothianidin-treated
99 weedy turf showed delayed weight gain and produced no new queens.²⁷ Recently, bees were
100 shown to become attracted to nicotinoid-containing nectar (sucrose solution containing im-
101 idacloprid or thiamethoxam) and that they cannot identify nicotinoids for avoidance.²⁸ Expo-
102 sure to imidacloprid, clothianidin and fipronil led to reductions in the proportion of active bees
103 in the hive, and initiated behaviours that reduced the efficiency of foraging flights. Also queens
104 of honey bees are negatively affected by neonicotinoids. Experimental exposure reduced their
105 reproductive anatomy and physiology.²⁹ Thus, it seems likely that bees living in farmland and
106 exposed to neonicotinoids suffer from sublethal effects and reduced survival.

107 Thus far, studies on adverse effects of neonicotinoids focused on orientation and forag-
108 ing behaviour and memory formation, as well as on population relevant traits, including colony
109 size and survival.^{29, 30} Surprisingly, molecular effects of neonicotinoids were very little regarded
110 (if at all), and thus poorly understood. Exposure to clothianidin led to a nAChR-dependent rap-
111 id mitochondrial depolarization in cultured neurons.³¹ Therefore, the aim of our study was to
112 assess molecular effects of four neonicotinoids to better understand the molecular basis of
113 their toxic actions in honey bees. We compared transcriptional effects on selected target
114 genes of nicotine with those of nitro- (clothianidin, imidacloprid and thiamethoxam) and cyano-
115 substituted (acetamiprid) neonicotinoids to identify common and compound-specific adverse
116 effect pathways. To this end, we focused on a series of different toxicologically relevant path-
117 ways, including neuronal signalling (*nicotinic acetylcholine receptor alpha 1 and 2 subunits*,
118 *nAChR α 1* and *nAChR α 2*), long-term memory formation (*creb* and *pka*), life-span (*vitellogenin*),
119 immune system response (*apidaecin* and *defensin-1*), and stress response (*catalase*). In addi-
120 tion to experimental laboratory exposures, we assessed effects of acetamiprid in the field. Our
121 study revealed for the first time important molecular effects at environmentally relevant con-
122 centrations that may explain physiological adverse effects on the bees and contribute to the
123 understanding of the decline of bee populations.

124

125 **Materials and Methods**

126

127 **Chemicals**

128 Nicotine (> 99% purity), acetamiprid, clothianidin, imidacloprid and thiametoxam (purities of all
129 > 99%) were purchased from Sigma–Aldrich (Buchs, Switzerland). Stock solutions for each
130 compound were prepared in DMSO and diluted into sucrose-solution (0.1% DMSO).

131

132 **Laboratory exposure experiments**

133 Foraging adult worker honeybees (*Apis mellifera carnica*) of mixed age were collected from
134 frames from an outdoor colony located in a rural site with no agricultural activity and pesticide
135 use in the Black Forest (Germany, GPS: N 47.7667, E 7.8333) from May to July 2014 and
136 2015. All of the bees used in the experiment were from the same colony. The colony had evi-
137 dence of *Varroa destructor* infestation and was treated with formic acid (August 2014) and
138 oxalic acid (December 2014). Honeybees were collected in small cylindrical plastic containers,
139 cold anaesthetised at 4°C for 60 minutes, and transferred into 16.5 × 11 × 6.5 cm³ PET bottles
140 with small holes in the lid for gas exchange. A larger hole of 2 mm served for holding an Ep-
141 pendorf microcentrifuge tube filled with 2 mL of a 20% sucrose solution containing nicotine and
142 individual neonicotinoids (concentrations Table S1) or 0.1 % DMSO as solvent control.

143 In a pilot study we investigated possible side effects of DMSO before we started the
144 real experiments. Two exposure experiments were conducted, the first one from April through
145 July 2014, the second one from April through July 2015. Each exposure experiment consisted
146 of three PET bottles with 10 bees per concentration and time point. Two of these three bottles
147 were used to isolate RNA and one of these three bottles was keep frozen as back-up. Each
148 exposure experiment was done twice. As three bees were pooled to obtain one RNA samples,
149 three technical replicates were obtained from one bottle. As RNA was isolated from two bot-
150 tles, there were two biological replicates consisting of 6 technical replicates per experiment. As
151 each exposure experiment was done twice, there were four biological replicates in total con-
152 sisting of 12 technical replicates.

153 Concentrations of neonicotinoids were selected on the basis of environmental realistic
154 levels in nectar, and higher doses were below LD₅₀ values. A summary of used concentrations
155 expressed as ng/bee and as ng/ml syrup is shown in Table 1. No compound related mortality
156 occurred during exposure. Ten randomly chosen adult worker bees were placed into each
157 bottle and bottles placed in an incubator (28°C, 60% humidity). Bees were fed ad libitum with
158 sucrose solutions containing the pesticides for 24, 48 and 72 h. Every 24 h, the 2 ml sucrose
159 solution was removed, the amount of sucrose solution taken up by the bees assessed, and
160 replaced by a new solution. The average amount of sucrose solution was 100 µL per bee
161 throughout all exposure experiments. Solvent control and exposed bees were removed at dif-
162 ferent exposure times (24, 48 and 72 h) and stored at -20°C until further analysis. In a pilot

163 study, we found no effects of the solvent control (DMSO) on expressional changes in the hon-
164 ey bees.

165

166 **Field study**

167 From May to July 2014, one bee hive (colony size around 50,000 bees) was placed into an
168 orchard, where plums were grown, near to the Swiss border in the Rhine valley (Blansingen,
169 GPS: N 47.689882, E 7.538758). Approximately 50-70 foraging bees were collected one day
170 before spraying neonicotinoid insecticide acetamiprid (brand name Mospilan®, water soluble
171 granulate with 200 g acetamiprid/kg with any adjuvants; 25 g Mospilan were dissolved in 100 L
172 water to get a 0.025 % solution) in the orchard, as well as one, three and seven days after
173 spraying. Pesticide application occurred on 16th of July 2014 employing a concentration of
174 0.125 kg/ha. Collected bees were transported to the laboratory and immediately frozen at -
175 20°C until RNA isolation.

176

177 **RNA isolation, reverse transcription, and quantitative (q)PCR**

178 The brain of frozen bees was removed in total by opening the cranium using a scalpel and
179 forceps. The thorax of frozen bees was removed by opening the chitin layer of the thorax using
180 a scalpel and removing the whole tissue in the thorax with forceps. The thorax was only re-
181 moved from bees of the study conducted 2014. Total RNA of three pooled bee brains or thorax
182 samples, respectively, was isolated using Trizol reagent according to the manufacturer's in-
183 structions. Per each condition, six biological replicates were isolated. 1000 ng RNA was re-
184 verse transcribed as described before.³² qPCR based on SYBR green fluorescence (SYBR
185 green PCR master mix; Roche) was performed as previously.³² Primer sequences were taken
186 from literature or newly designed using the NCBI primer BLAST tool and sequences of used
187 primers are given in Table S1. For all performed analysis *ribosomal protein L32 (rpl32)* was
188 used as house-keeping gene for normalisation due to constant expression in bees, similar as
189 in other systems.³² All samples showed constant *rpl32* expression (Figure S1). Alterations of
190 mRNA abundance in neonicotinoid exposed samples were always compared against the sol-
191 vent (DMSO) control samples to determine the effect of pesticides.

192

193 **Data processing and statistical analysis**

194 Heat maps of expressional changes were designed by importing analysed qPCR data into
195 MEV 4.9 (Multi Experiment Viewer) software. Differences between treatments were assessed
196 by one way ANOVA followed by a Bonferroni's multiple comparison test to compare treatment
197 means with respective controls. Results are given as means ± standard error of means. Differ-
198 ences were considered statistically significant with one asterisk at $0.05 > p > 0.01$, two aster-
199 isks at $0.01 > p > 0.001$ and three asterisks at $0.001 > p > 0.0001$. All statistical data of the

200 one way ANOVA are provided in Table S2. Correlation between field data and laboratory data
201 were analysed by linear regression.

202
203
204

205 Results

206

207 First, we evaluated whether transcriptional changes induced by thiamethoxam were similar in
208 brain and thorax of honey bees. To this end transcript of six genes were compared in bees
209 exposed to 0.1, 1, 2.5 and 5 ng/bee thiamethoxam for 24, 48 and 72 h (Fig. S2). We found a
210 significant increase of *AChR α 1*, *AChR α 2*, *vitellogenin* and *catalase* transcripts in both tissues.
211 *Apidaecin* was significantly up-regulated in the brain only. As both tissues showed almost simi-
212 lar transcriptional alterations, subsequent analyses were focused solely on the brain as target
213 organ of neonicotinoids although extrapolation from one tissue to another is difficult.

214

215 We then compared effects of nicotine to those of the four neonicotinoids acetamiprid,
216 clothianidin, imidacloprid and thiamethoxam, focusing on different toxicological pathways in
217 the brain. Some of the effects of nicotine and all three neonicotinoids are shared, while some
218 compound-specific patterns occurred in some of the target genes.

218

219 Transcriptional changes on target genes and pathways

220

221 We assessed the expression of eight selected target genes in the brain of honey bee workers
222 after exposure for 24, 48 and 72 h to nicotine and the four neonicotinoids. Nicotine led to in-
223 duction of the *nAChR α 1* transcript after 48 h and after 72 h, acetamiprid, clothianidin and im-
224 idacloprid after 72 h and thiamethoxam after 48 h and 72 h (Fig 1A). Significant induction of
225 *nAChR α 2* was detected after exposure to imidacloprid and thiamethoxam after 48 h, but not
226 after nicotine, acetamiprid and clothianidin exposure (Fig. 1B). The strongest significance for
227 both transcripts was found for thiamethoxam with $F(14, 165) = 6.9$ and 5.327 (Table S2).

227

228 We observed strongest and dose-related increases of the *vitellogenin* transcript after
229 exposure to thiamethoxam in all concentrations after 48 h, and in the two highest concentra-
230 tions after 72 h exposure (Fig. 2). While nicotine did not induce this transcript, significant in-
231 creases occurred for acetamiprid and imidacloprid after 24 h and after 48 h (Fig. 2). Cloth-
232 ianidin induced the *vitellogenin* transcript after 24 h (Fig. 2). Thiamethoxam showed the
233 strongest significance with $F(14, 165) = 6.745$ (Table S2).

233

234 Nicotine and thiamethoxam did not alter the amount of the *pka* transcript, while acet-
235 amiprid and imidacloprid lowered its amount after 72 h (Fig. 3). Clothianidin led to decrease of
236 the *pka* transcript at all time points, with strongest effect after 48 h (Fig. 3). Whereas acetam-
237 iprid and thiamethoxam did not change the *creb* transcript, imidacloprid led to a decrease after
48 and 72 h, and nicotine led to a decrease after all time points. In contrast, clothianidin led to

238 an increase after 24 h (Fig. 4). Clothianidin showed for both transcripts the strongest significant
239 changes with $F(14, 165) = 5.261$ for *pka* and 5.906 for *creb* (Table S2).

240 Nicotine had no effect on the amount of *apidaecin* and *defensin-1* transcripts (Fig. 5,
241 Fig. S3). Acetamiprid and imidacloprid lowered the amount of *apidaecin* transcript after 72 h,
242 clothianidin after 24 h, while thiamethoxam led to an increase after 48 h (Fig. 5). The *defensin-*
243 *1* transcript was increased at different exposure times upon acetamiprid, clothianidin, im-
244 idacloprid and thiamethoxam exposure (Fig. S3). In case of the *catalase* transcript, acetam-
245 iprid and thiamethoxam led to an increase after 48 h, and imidacloprid led to concentration-
246 dependent alterations, whereas nicotine and clothianidin had no effect on the expression of
247 this transcript (Fig. S4). Changes for these three transcripts showed only weak significance
248 (Table S2). Thus, the most prominent transcriptional alterations were found for *AChR α 1* and 2,
249 *creb*, *pka* and *vitellogenin*. In particular the up-regulation of *vitellogenin* is specific for neonicoti-
250 noids, but not for nicotine.

251

252 Overall change and transcripts profiles of nicotine and neonicotinoids

253 To compare and summarize overall changes in gene expression, a heat map analysis was
254 performed. Thereby, changes during different exposure times are visible for each transcript.
255 The strongest transcriptional down-regulations were found for acetamiprid and clothianidin,
256 and the strongest transcriptional induction was found for thiamethoxam (Fig. 6A). The two low
257 concentrations of thiamethoxam and clothianidin showed very similar expression profiles. The
258 same holds true for the two high thiamethoxam concentrations (Fig. S5). The low imidacloprid
259 concentration showed a similar expression pattern as thiamethoxam and the middle and high
260 concentrations induced a similar pattern as acetamiprid (Fig. S5). Acetamiprid, imidacloprid
261 and thiamethoxam build one group with similar transcriptional alterations, which was distinct
262 from that of clothianidin and of nicotine (Fig S5). Expression profiles of different transcripts
263 show that *nAChR α 1* (24 h), *nAChR α 2* (24 h), *nAChR α 1* (72 h), *nAChR α 2* (72 h), *defensin-1*
264 (72 h) and *pka* (72 h) built one group. A second group consists of *nAChR α 1* (48 h), *nAChR α 2*
265 (48 h), *apidaecin* (72 h), *catalase* (48 and 72 h), *defensin-1* (48 h), *pka* (48 h) and *vitellogenin*
266 (72). And the transcripts of *catalase* (24 h), *creb* (24, 48 and 72 h) and *vitellogenin* (24) built a
267 third group. These three groups built again one group with similar expression pattern. Very
268 different expression profiles were found for *apidaecin* (24 h), *defensin-1* (24 h), *pka* (24 h) and
269 *vitellogenin* (48 h) (Fig. S5).

270 This heat map analysis indicates that the transcription profiles of nicotine and neonicoti-
271 noids induced different expression patterns, although binding to the same receptor. The low-
272 est effect concentrations for clothianidin, imidacloprid and thiamethoxam in our study were
273 below reported residue concentrations in pollen and nectar (Fig. 6B).^{27, 28, 29} This indicates that

274 our observed transcriptional effects occur at environmentally realistic concentrations in case of
275 clothianidin, imidacloprid and thiamethoxam.

276

277 **Effects of acetamiprid in the field study**

278 Our field study, where acetamiprid (Mospilan®) was applied, allowed a direct comparison to
279 the experimental laboratory exposures. A strong increase in transcript abundance occurred for
280 *nAChR α 1* and *nAChR α 2*, *vitellogenin* and *catalase* one day after application of the pesticide in
281 bees, while *Apidaecin*, *creb*, *defensin-1* and *pka* showed no significant changes (Fig. 7A). The
282 detected alterations in gene expression were highly significant with $F(31, 160) = 16.29$ (Table
283 S2). We compared transcriptional changes in bees in the field study after one day to those
284 found in bees after 24 h of laboratory exposures to low acetamiprid concentrations. The heat
285 map shown in Fig. 7B suggests a similar, but not identical pattern. The transcripts of *nA-*
286 *ChR α 1*, *nAChR α 2* and *catalase* reacted very similarly. Along with *vitellogenin* and *apidaecin*,
287 they built one cluster. *Creb* and *defensin-1* also showed similar expression profiles, while *pka*
288 was different (Fig. 7B). In higher acetamiprid concentrations the transcripts pattern showed
289 larger deviations to the field exposure. The correlation between transcript levels in the low and
290 high acetamiprid concentrations and the field-study on day one after pesticide application was
291 not significant (Fig. 7C). Thus, the field exposure showed much stronger transcriptional
292 changes than laboratory exposures, but some similarities in expression patterns to the exper-
293 imental acetamiprid exposures occurred.

294

295

296

297 **Discussion**

298 Here we showed for the first time significant molecular effects of nicotine and neonicotinoids
299 in the brain of honey bee workers in experimental laboratory and field exposures. Tran-
300 scripts of *nAChR α 1* and *nAChR α 2* were induced by nicotine and neonicotinoids, while *vitello-*
301 *genin* was induced by neonicotinoids only. In addition to affecting the neuronal system, neonicotinoids
302 led to expressional changes of immune system related genes, suggesting adverse
303 effects on brain function and immune system defence. Effects were stronger for the three
304 more toxic neonicotinoids, clothianidin, imidacloprid and thiamethoxam than for the less toxic
305 acetamiprid. Our data also indicate that clothianidin as reactive metabolite of thiamethoxam
306 resulted in similar but faster and more significant transcriptional alterations; they occurred at a
307 shorter exposure time than those of thiamethoxam. We also showed that similar molecular
308 effects of acetamiprid occurred in field exposures as in the laboratory. Effects observed in the
309 field exposure were stronger but transient. A schematic overview of the effects of neonicotinoids
310 is given in Figure 8. They are characterized by induction of *vitellogenin*, down-

311 regulation of *apidaecin* and *pka*. We hypothesize that these molecular effects may represent a
312 molecular basis for the previously reported physiological and behavioural effects.

313 The effect concentrations found in our study are much lower than reported LD₅₀ values
314 of clothianidin, imidacloprid and thiamethoxam although different LD₅₀ values can be found for
315 these three neonicotinoids in the literature. Laurino et al. determined LD₅₀ values of 4.48
316 ng/bee at 24 h, 4.32 ng/bee at 48 h, and 4.21 ng/bee at 72 h for clothianidin; 183.78 ng/bee
317 at 24 h, 104.12 ng/bee at 48 h and 72.94 ng/bee at 72 h for imidacloprid, and 3.55 ng/bee at
318 24 h, 3.35 ng/bee at 48 h and 2.88 ng/bee at 72 for thiamethoxam.³³ 2013, different LD₅₀ val-
319 ues were determined by the same group for clothianidin, imidacloprid and thiamethoxam.³⁴ The
320 LD₅₀ value of acetamiprid was 14.5 µg/bee.¹⁵ In our study, we did not observe acute toxicity at
321 any of the tested concentrations, so our observed effects seem not off-target effects. The ob-
322 served effects concentrations are also much lower than known IC₅₀ values of insect nAChR
323 that are in the mg/mL range.^{36, 37} This indicates that the observed effects at low concentrations
324 are due to specific interactions of neonicotinoids with nAChRs. Nevertheless, at higher con-
325 centrations the observed effects could also be due to activation of muscarinic AChR (second-
326 ary effect). To distinguish contributions between nAChR and muscarinic AchR activation, ex-
327 periments with nAChR antagonists would be needed. However, based on the similarity of ef-
328 fects at low and high neonicotinoid concentrations, contributions of muscarinic AChR seem to
329 be of minor importance.

330 We found that nicotine and neonicotinoids led to up-regulation of *nAChR* transcripts in
331 experimental exposures and in our field study in foraging honey bees. Our data confirm those
332 found for nicotine.⁴⁰ Nicotine also affected short-term memory.⁴¹ Thus, neonicotinoids have the
333 same effects as nicotine on the expression of *nAChRs*. Upregulation of *nAChRs* in response
334 to neonicotinoids can be regarded as a compensation reaction to the functional loss of these
335 receptors, thus suggesting auto-regulation. Chronic administration of nicotine triggered the up-
336 regulation of *nAChRs* in rats.⁴² Alterations in neuronal signalling can have pronounced effects,
337 as indicated by exposure of honey bees to 3.8 ng/bee thiamethoxam, which induced locomotor
338 deficits.⁴³ Exposure of honey bees to 0.34 ng/bee clothianidin or 0.40 ng/bee imidacloprid or
339 0.48 ng/bee thiamethoxam induced adverse effects on basic motor function.⁴⁴

340 cAMP-dependent kinase (PKA) regulates multiple cellular processes, including the
341 formation of long-term memory in honey bees.⁴⁵ Once activated, PKA induces the phosphory-
342 lation of cAMP response element binding protein (Creb). Phosphorylated Creb acts as an ac-
343 tive transcription factor and induces the expression of Creb target genes that are thought to
344 contribute to the formation of long-term synaptic plasticity which is important for long-term
345 memory consolidation.⁴⁶ In addition, Creb is known to play a central role in the formation and
346 consolidation of memory.^{47, 48, 49} Creb is also involved in the visual and olfactory learning in
347 bees.⁵⁰ The observed down regulation of *creb* upon neonicotinoid exposure in our experiments

348 could be involved in the adverse effects on memory formation. Inhibition of Creb in rats was
349 responsible for the nicotine-induced impairment of hippocampal plasticity.⁴⁸ Likewise, down-
350 regulation of *creb* in honey bees could lead to reduction of plasticity of the neuronal system.
351 We found a down-regulation of *creb* by nicotine and imidacloprid, and a down-regulation of
352 *pka* by acetamiprid, clothianidin and imidacloprid. This could partly explain the negative effects
353 of neonicotinoids on long-term memory formation and learning. It is known that 6.12 ng/bee
354 imidacloprid and 0.69 pg/bee thiamethoxam impaired short-term olfactory memory in foraging
355 honey bees.⁵¹

356 Vitellogenin is a female-specific glucolipoprotein produced under hormonal control and
357 produced by the reproductive queen, but also by worker bees, particularly hive bees. When
358 hive bees develop into foragers, vitellogenin production is ceased. Vitellogenin is predominant-
359 ly found in the hemolymph of queens, hive bees and winter bees. The levels are highest in the
360 longest-lived winter bee workers (up to 60–90 µg/µl hemolymph), and lowest in short-lived
361 foragers.⁵² In addition to reproductive function (queen) vitellogenin can be characterized as a
362 protein being used for different metabolic purposes and it also shields cells from oxidative
363 damage and protects against oxidative stress.⁵² The protectant role and its increase with ag-
364 ing, led to the hypothesis that vitellogenin serves as a regulator of honey bees lifespan.^{52, 53}
365 Furthermore, vitellogenin can suppress juvenile hormone in worker bees,⁵⁴ which acts as pro-
366 aging hormone⁵⁵ and promotes foraging. Down-regulation of vitellogenin by RNA interference
367 led to increased levels of juvenile hormone and accelerated transition to the short-lived forager
368 stage.⁵⁵ A drop in vitellogenin titre regulates the behavioural shift to foraging, and increases
369 the juvenile hormone level. Juvenile hormone titres of hive bees, which are not foraging, were
370 10-100 ng/ml, and of foragers, 200-600 ng/ml.^{56, 57} RNAi-mediated silencing of vitellogenin
371 gene function turns honey bee workers into extremely precocious foragers. This suggests that
372 vitellogenin is a primary switch signal.⁵⁸ An increased expression of vitellogenin was observed
373 in non-foraging worker bees in queen-less colonies.⁵⁹

374 In addition, vitellogenin regulates foraging performance in worker bees. With high lev-
375 els, workers search more for pollen, and with low levels, for nectar.⁶⁰ Exposure of honey bees
376 to clothianidin and imidacloprid in the low ng/bee range reduced foraging activity.⁶¹ So far, the
377 decrease in foraging activity upon neonicotinoid exposure was linked to disturbed memory
378 formation. Neonicotinoids clothianidin, imidacloprid and thiamethoxam were thought to either
379 block the retrieval of exploratory navigation memory, or altered this form of navigation
380 memory.⁶² The up-regulation of vitellogenin after neonicotinoid exposure observed in our study
381 could be another explanation for the previously shown reduced foraging activity of honey bee
382 workers after neonicotinoid exposure.⁶¹ Most notably, the strong induction of vitellogenin de-
383 tected in bees exposed to acetamiprid in our field study may be an explanation for transient
384 negative effects of neonicotinoids on foraging activity of honey bees. The importance of *vitel-*

385 *logenin* as a biomarker for neonicotinoid exposure is supported by comparison with nicotine;
386 *vitellogenin* was the only gene studied that was strongly up-regulated upon experimental neon-
387 icotinoid exposure. Strong increase was also found in our field study. Therefore, the expres-
388 sion of *vitellogenin* may serve as a good biomarker candidate for the exposure of honey bees
389 to neonicotinoids. Further analysis addressing this in more depth, as well as studies on the
390 vitellogenin protein levels would help to understand this important effect.

391 Insect immunity shows many parallels to the innate immune response of vertebrates,
392 involving a diverse set of actions including the secretion of antimicrobial peptides, phagocyto-
393 sis, melanisation and the enzymatic degradation of pathogens.⁶³ Apidaecin and defensin-1 are
394 members of the Toll like receptor pathway and function as antimicrobial effectors.⁶¹ Injection of
395 *E. coli* or the bee parasite *Paenibacillus larvae* into adult worker bees lead to the strong induc-
396 tion *apidaecin* and *defensin-1* transcripts.⁶⁴ This shows the important role of both proteins to
397 fight against pathogens. Exposure to field realistic concentration of honey bees to imidacloprid
398 decreased hemocyte density, encapsulation response and antimicrobial activity.⁶⁵ Adverse
399 effects of nicotine on the immune system were previously reported.⁶⁶ Decreased transcripts of
400 *apidaecin* found in our experiments upon exposure to acetamiprid, clothianidin and imidaclo-
401 prid may be implicated in negatively effects on the immune system. In contrast to *apidaecin*,
402 *defensin-1* was upregulated upon neonicotinoid exposure (Fig. S3). Defensin is produced as
403 antimicrobial peptide upon infection with pathogens.⁶⁷ There is a negative correlation between
404 the expression of defensin-1 and the expression of storage proteins like apoLp-III,⁶⁸ thus neon-
405 icotinoid-related up-regulation of defensin-1 may affect this process. Interestingly, nicotine
406 and neonicotinoid-induced transcriptional changes did not show clear dose-effect relation-
407 ships, but can rather be regarded as an all or nothing reaction. Only clothianidin and thiameth-
408 oxam showed some dose-dependent effects. Moreover, sometimes strongest reaction oc-
409 curred in the lowest dose. This is in line with previous data on sub-chronic toxicity of neonic-
410 otinoids.^{69,70} These non-linear patterns could be due to the complex interaction between nAChR
411 binding of neonicotinoids and gene expression regulation that can generate non-classical dose
412 response relationships.⁷¹

413 Especially for the three highly toxic neonicotinoids, clothianidin, imidacloprid and thia-
414 methoxam, effects occurred at very low ng/bee (or low ng/mL) levels. These concentrations
415 are environmentally relevant. Reported maximal concentrations of imidacloprid and thiameth-
416 oxam in wild flowers collected adjacent to recently planted crop fields were 48 ng/g and 256
417 ng/g.⁷² In pollen of squash, imidacloprid and thiamethoxam concentrations were up to 28 ppb
418 (28 ng/mg) and 35 ppb (35 ng/mg), respectively. The maximal found imidacloprid and thia-
419 methoxam concentrations in nectar of squash were 14 ppb (14 ng/ml) and 20 ppb (20 ng/ml),
420 respectively.³⁵ In pollen of rape flowers, mean clothianidin and thiamethoxam concentrations
421 were 2.27 ng/g and 3.26 ng/g, respectively. In pollen of wild flowers grown at the margins of

oilseed rape fields, mean thiamethoxam concentrations were 14.81 ng/g.⁷³ In addition to nectar and pollen, puddle water is a source for neonicotinoid uptake; thiamethoxam concentrations of up to 63.4 $\mu\text{g/l}$ (63.4 ng/ml) were found.⁷⁴ Thus, environmental concentrations of clothianidin, imidacloprid and thiamethoxam are in the same range of the here reported concentrations that induced alterations in gene expression. Clothianidin and imidacloprid induced the expression of *vitellogenin* at 3 ng/ml and thiamethoxam at 0.1 ng/ml. Therefore, the observed effects in our study occurred at environmental realistic concentrations. The fact that our field study revealed strong transcriptional changes that were similar to the laboratory study with acetamiprid, namely upregulation of *AChRs*, *vitellogenin* and *catalase*, strengthens the implication that our experimental laboratory data can be extrapolated to the field. However, our study focused on the transcriptional level and it is currently not clear, how the transcriptional alterations translate to the protein level and to physiological effects. This limitation should be addressed in forthcoming studies by analyse on the protein level and by linking molecular with physiological effects.

In conclusion, here we report for the first time molecular effects on key neuronal processes in the brain of honey bees exposed to four neonicotinoids and compare it to those of nicotine. Most prominent transcriptional alterations occurred for *AChR α 1* and 2, *creb*, *pka* and *vitellogenin*, which seems selective for the neonicotinoids (but not for nicotine). Clothianidin and thiamethoxam showed dose-response effects for certain transcripts (clothianidin: *AChR α 2*, *creb* and *vitellogenin*; thiamethoxam: *AChR α 1*, *AChR α 2* and *vitellogenin*), confirming that they are compound-related. The observed transcriptional effects may represent a basis for previously reported adverse effects of neonicotinoids on the immune system, foraging behavior and memory in bees, and provide preliminary insights into the molecular basis of such physiological effects. Particularly the induction of *vitellogenin* may have physiological implications and can serve as potential biomarker for neonicotinoid exposure. Bees exposed to neonicotinoids showed reduced foraging activity and high vitellogenin levels (normally found in winter bees). Therefore, the observed overexpression of the *vitellogenin* transcript could be one of the molecular bases for the alteration in foraging activity upon neonicotinoid exposure. The alterations in *creb* and *pka* expression may represent a molecular basis for the observed negative effects of neonicotinoids on honey bee memory formation. Therefore, our work may provide some first hints on the molecular basis of the observed physiological changes of honey bees upon neonicotinoid exposure. How these transcriptional alterations will translate to adverse effects in bees living in neonicotinoid contaminated environments should be addressed in forthcoming studies. Our data clearly show adverse effects of neonicotinoids at environmental concentrations and help to understand the potential contribution of neonicotinoids to the decline of bee populations.

458

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464

465 Associated content:

466 Supporting information

467 Information about primer sequences used for quantitative real-time PCR analysis and statisti-
468 cal values of the ANOVA analysis (Tables S1 and S2). Expression of reference gene *rp/32*
469 (Figure S1), results of the laboratory study performed in 2014 (Figure S2), abundance of tran-
470 script *defensin-1* after exposure to nicotine and neonicotinoids (Figure S3), abundance of tran-
471 script *catalase* after exposure to nicotine and neonicotinoids (Figure S4), and overall gene
472 expression analysis of the laboratory study (Figure S5).

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- 759

760 Table 1: Concentration of nicotine and the four neonicotinoids used in the present study.

761

| Compound | Concentration (ng/bee) | Concentration (ng/ml sugar syrup) |
|--------------|------------------------|-----------------------------------|
| Nicotine | 486, 4860 and 48600 | 4860, 48600 and 486000 |
| Acetamiprid | 8; 80; 800 and 8000 | 80; 800; 8000 and 80000 |
| Clothianidin | 0.03; 0.3; 1.5 and 3 | 0.3; 3; 15 and 30 |
| Imidacloprid | 0.3; 3 and 30 | 3; 30 and 300 |
| Thiamethoxam | 0.01; 0.05; 0.1 and 1 | 0.1; 0.5; 1 and 10 |

762

763 **Figure legends**

764

765 **Figure 1**

766 Abundance of transcripts of *nAChR α 1* (A) and *nAChR α 2* (B) in the brain of honeybees follow-
767 ing exposure to different concentrations of nicotine, acetamiprid, clothianidin, imidacloprid and
768 thiamethoxam for 24 h (vertical strips), 48 h (squares) and 72 h (horizontal strips). Shown are
769 the results of four biological replicates per concentration and time-point. Significant differences
770 with p-value of ≤ 0.05 are marked with asterisks.

771

772

773 **Figure 2**

774 Abundance of transcripts of *vitellogenin* in the brain of honeybees following exposure to differ-
775 ent concentrations of nicotine, acetamiprid, clothianidin, imidacloprid and thiamethoxam for 24
776 h (vertical strips), 48 h (squares) and 72 h (horizontal strips). Shown are the results of four
777 biological replicates per concentration and time-point. Significant differences with p-value of \leq
778 0.05 are marked with asterisks.

779

780

781 **Figure 3**

782 Abundance of transcripts of *pka* in the brain of honeybees following exposure to different con-
783 centrations of nicotine, acetamiprid, clothianidin, imidacloprid and thiamethoxam for 24 h (ver-
784 tical strips), 48 h (squares) and 72 h (horizontal strips). Shown are the results of four biological
785 replicates per concentration and time-point. Significant differences with p-value of ≤ 0.05 are
786 marked with asterisks.

787

788 **Figure 4**

789 Abundance of transcript of *creb* in the brain of honeybees following exposure to different con-
790 centrations of nicotine, acetamiprid, clothianidin, imidacloprid and thiamethoxam for 24 h (ver-
791 tical strips), 48 h (squares) and 72 h (horizontal strips). Shown are the results of four biological
792 replicates per concentration and timepoint. Significant differences with p-value of ≤ 0.05 are
793 marked with asterisks.

794

795 **Figure 5**

796 Abundance of transcripts of *apidaecin* in the brain of honeybees following exposure to different
797 concentrations of nicotine, acetamiprid, clothianidin, imidacloprid and thiamethoxam for 24 h
798 (vertical strips), 48 h (squares) and 72 h (horizontal strips). Shown are the results of four bio-
799 logical replicates per concentration and time-point. Significant differences with p-value of \leq
800 0.05 are marked with asterisks.

801 Figure 6

802 Expression profiles in the brain of honey bees following exposure to nicotine, acetamiprid, im-
803 idacloprid and thiamethoxam for different times and concentrations. (A) Heat map of all ob-
804 tained transcriptional alterations for each compound and concentration shown for different
805 exposure times of 24, 48 and 72 h. (B) Comparison of effect levels found in the present study
806 (only significant effects with an F-statistic above 3) with reported pollen and nectar concentra-
807 tions of acetamiprid, clothianidin, imidacloprid and thiamethoxam according to ^{35, 38, 39} Red line
808 shows the lowest observed effect concentrations (LOEL) observed in the present study.

809

810

811 Figure 7

812 (A) Abundance of transcripts of eight selected genes in the brain of foraging honey bees one
813 day (vertical strips), three days (squares) and seven days (horizontal strips) after exposure to
814 acetamiprid (Mospilon®). (B) Comparison of overall alterations in gene expression after exper-
815 imental exposure to acetamiprid for 24 h and field exposure to acetamiprid (Mospilon®) for
816 one day in the field study by heat map analysis. Shown are the results of six pooled samples
817 per concentration and time-point. Significant differences with p-value of ≤ 0.05 are marked by
818 asterisks. (C) Correlation between alteration in gene expression (fold-changes) for determined
819 transcripts between field-study-day 1 and the lowest (left, r^2 : 0.32) and the highest (right, r^2 :
820 0.25) acetamiprid concentration after experimental 24 h exposure.

821

822

823 Figure 8

824 Schematic overview of neonicotinoid-induced molecular effects in honey bees on the transcrip-
825 tional level, and associated physiological and behavioural consequences. Red arrows point
826 out up- and down-regulation of the investigated transcripts.

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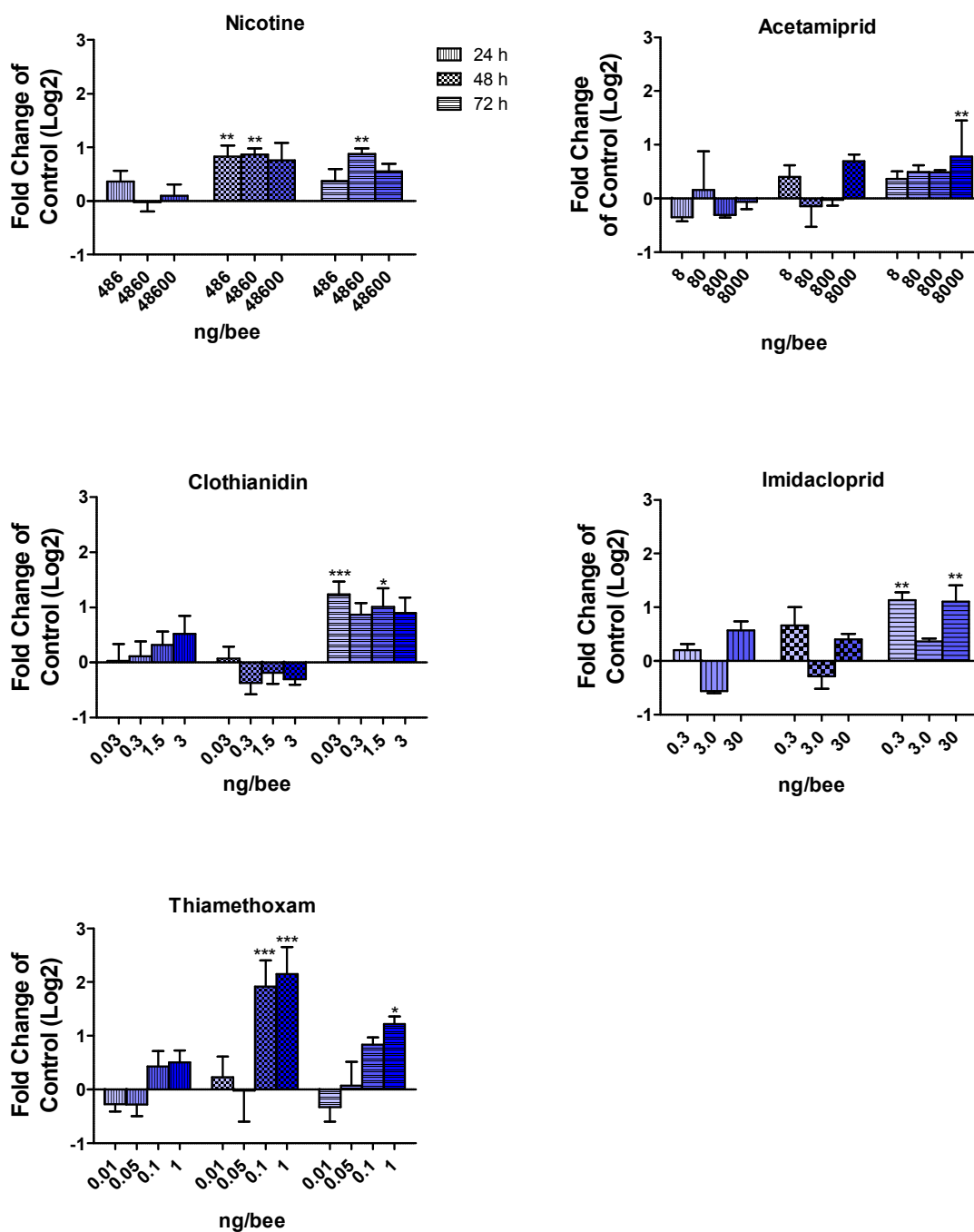
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842 Fig. 1:

843 **A**



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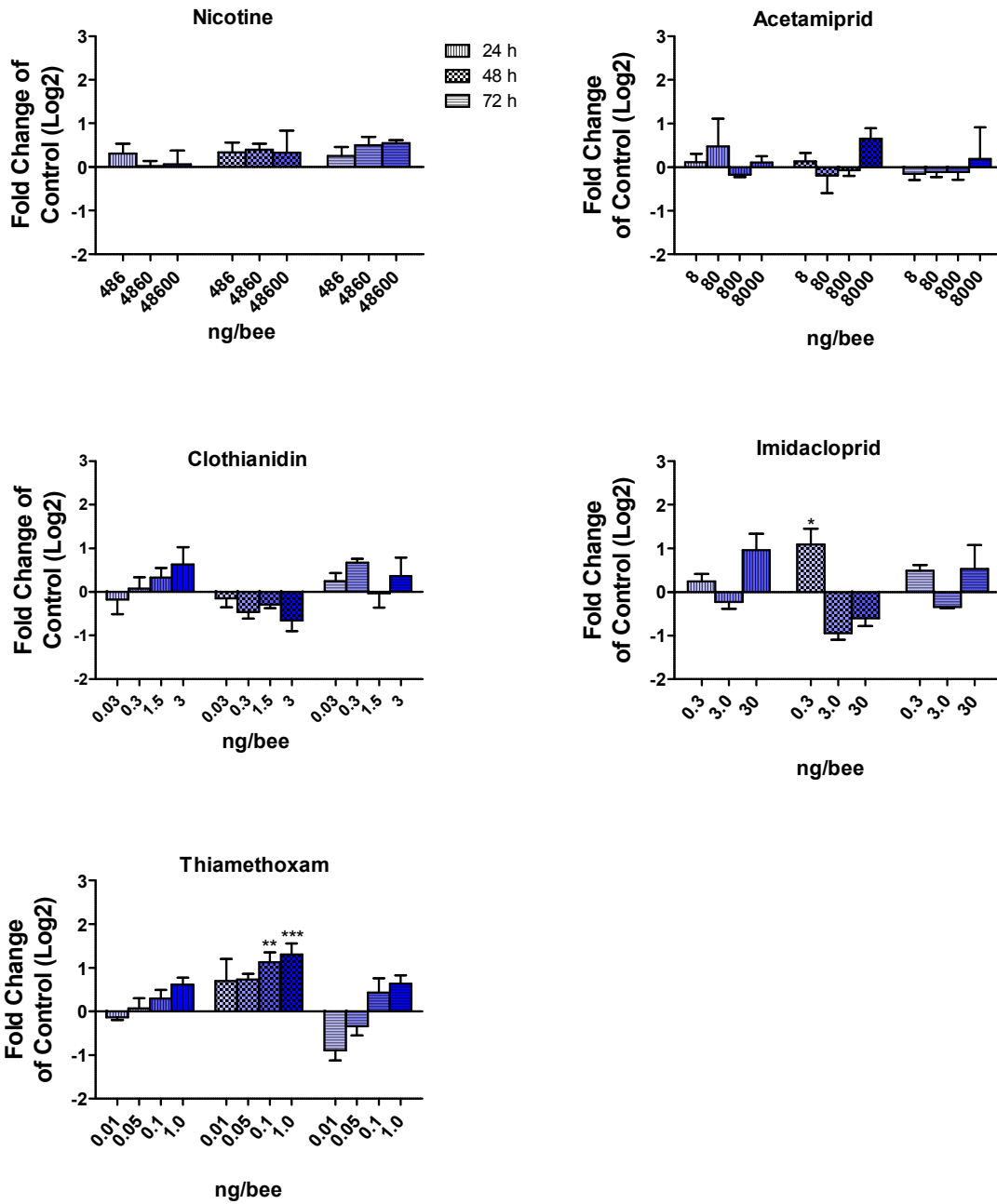
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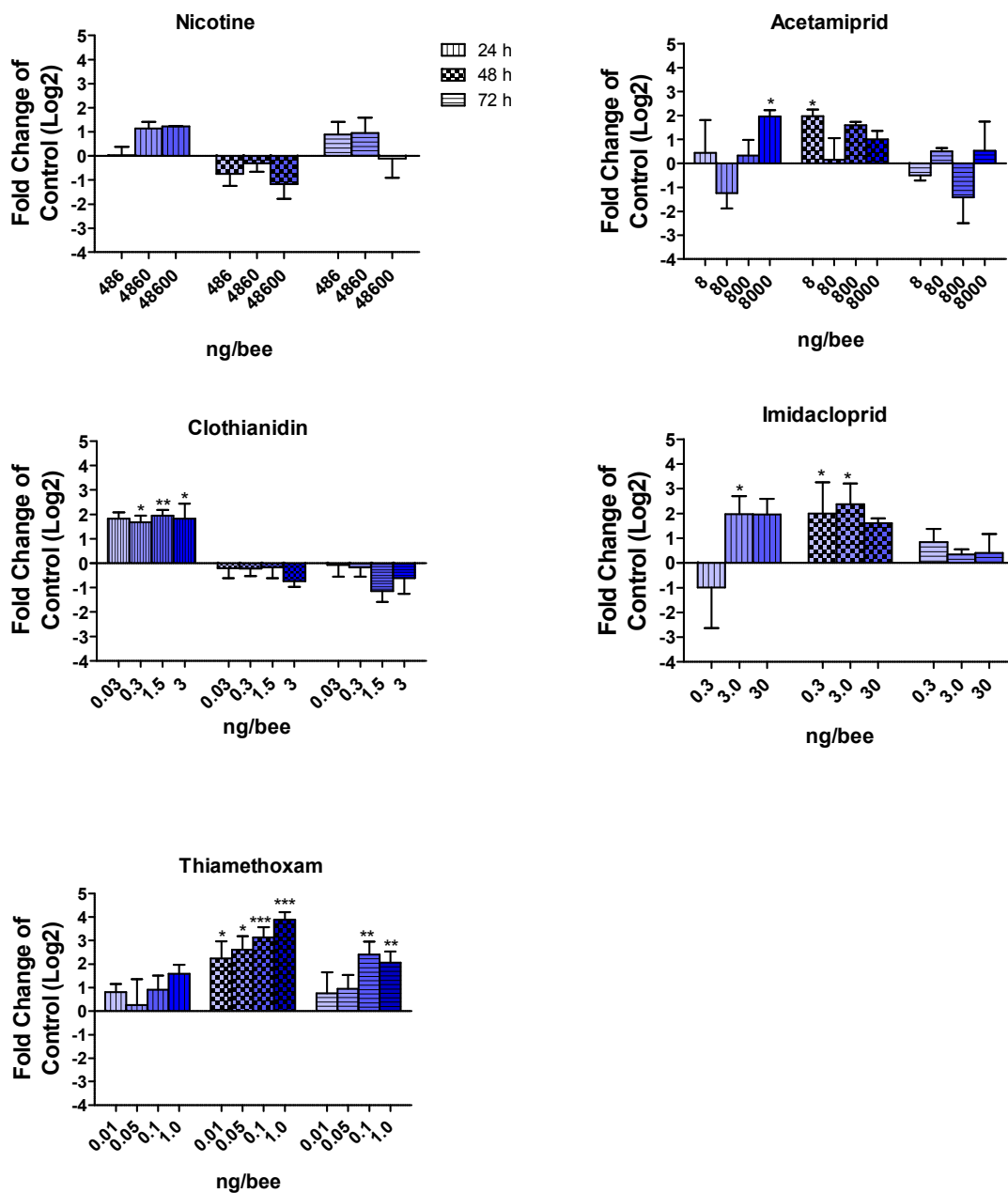
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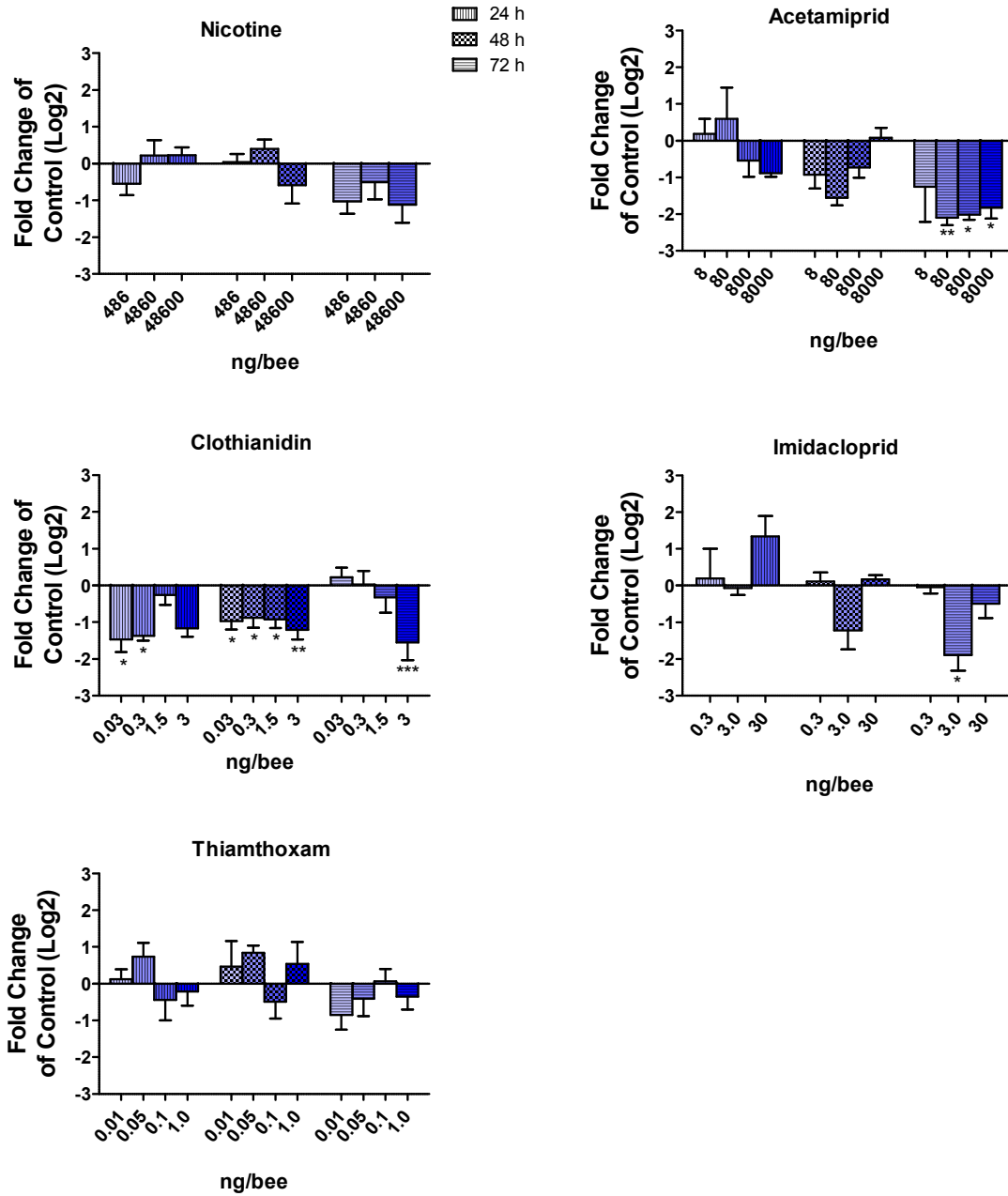
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857 Fig. 2:
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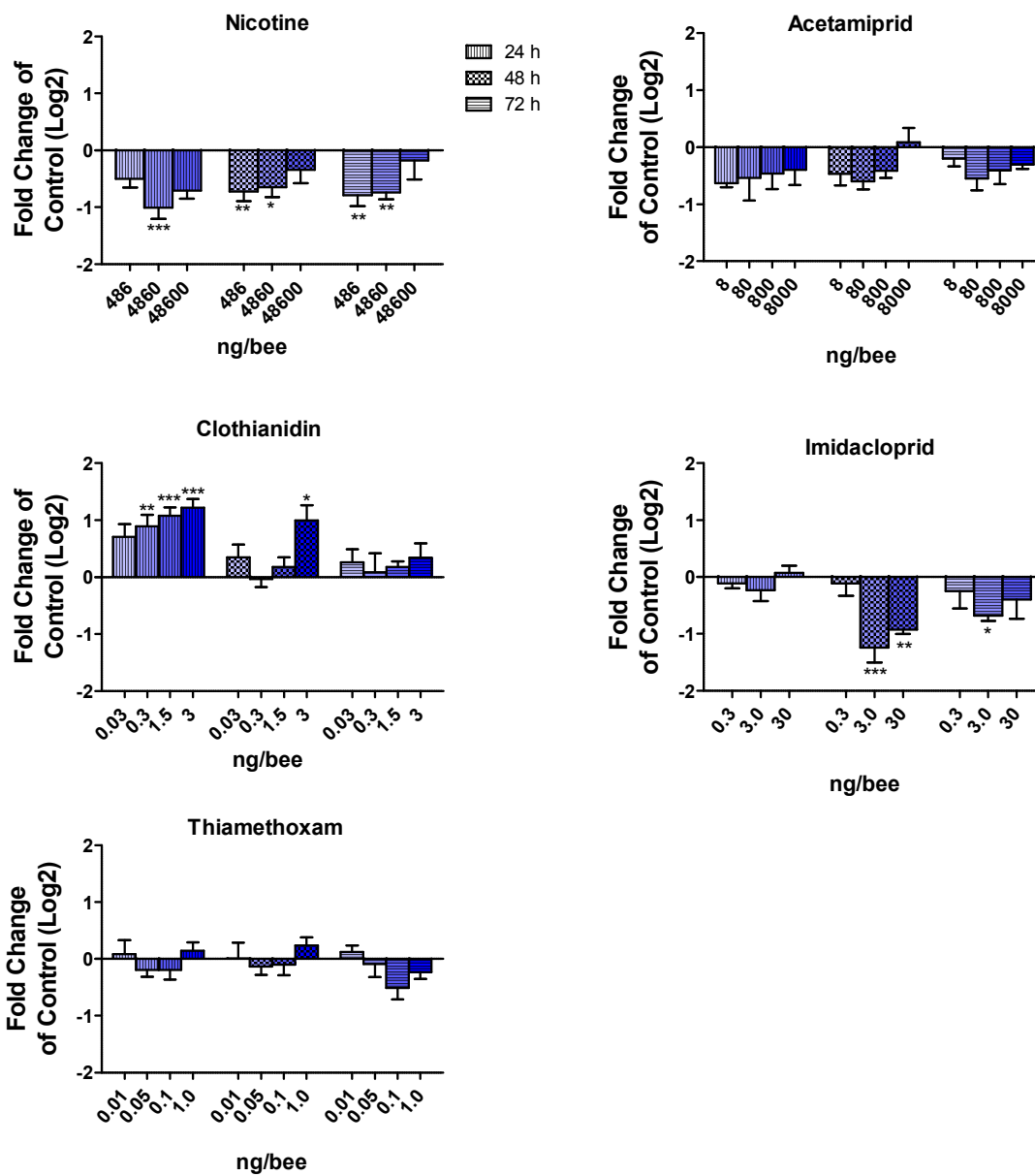
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862 Fig. 3:



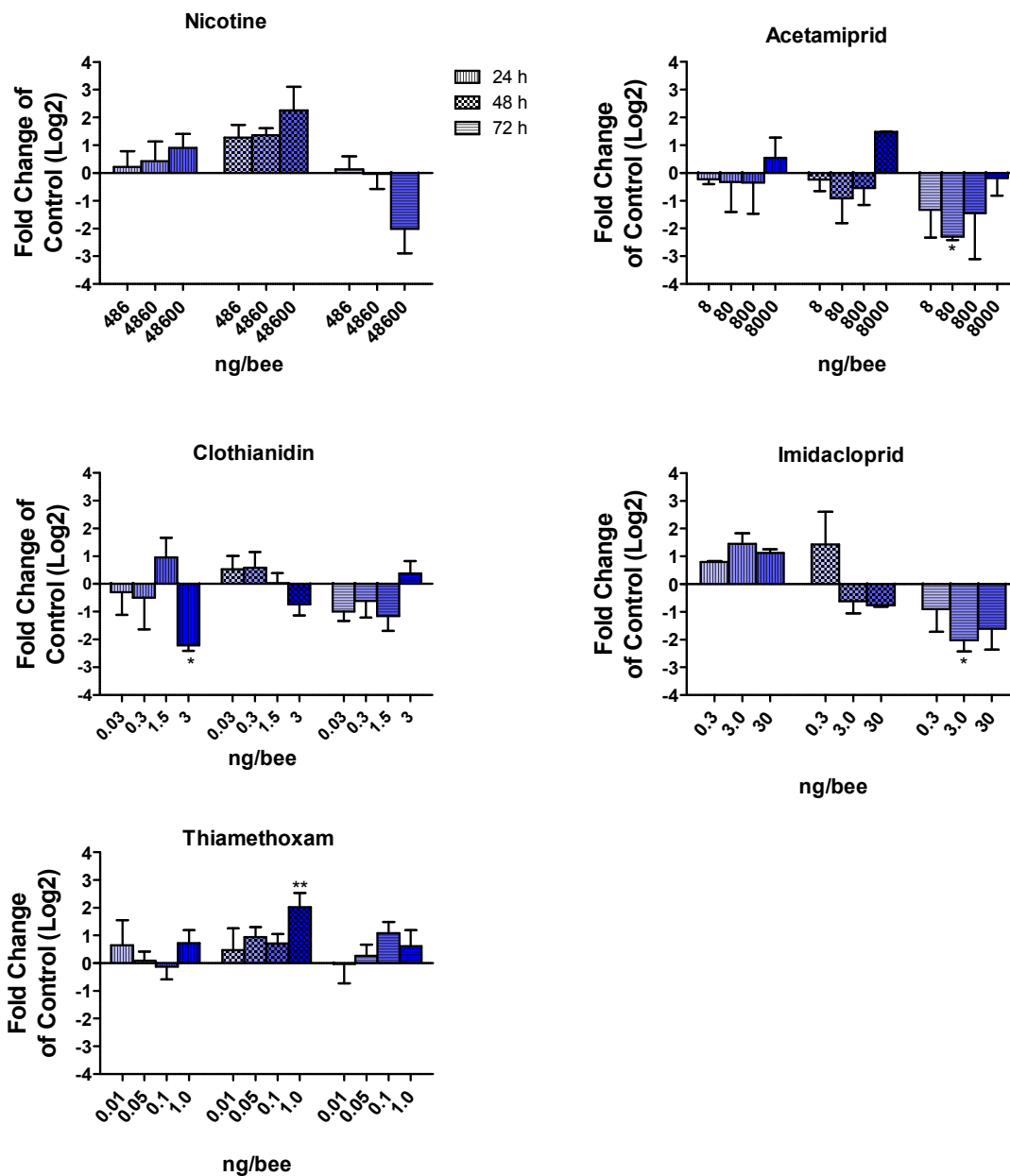
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866 Fig. 4:
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870 Fig. 5:



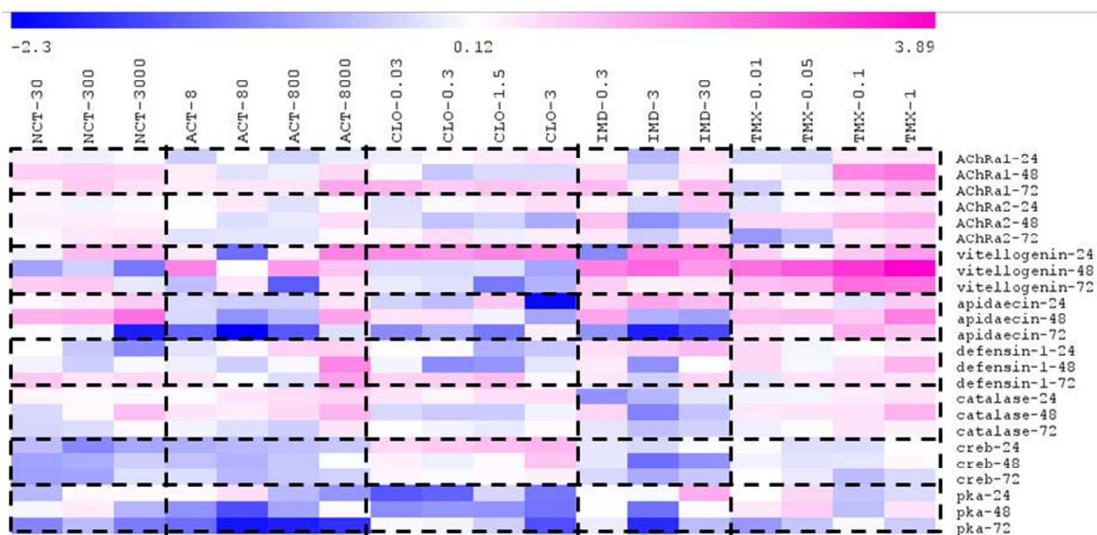
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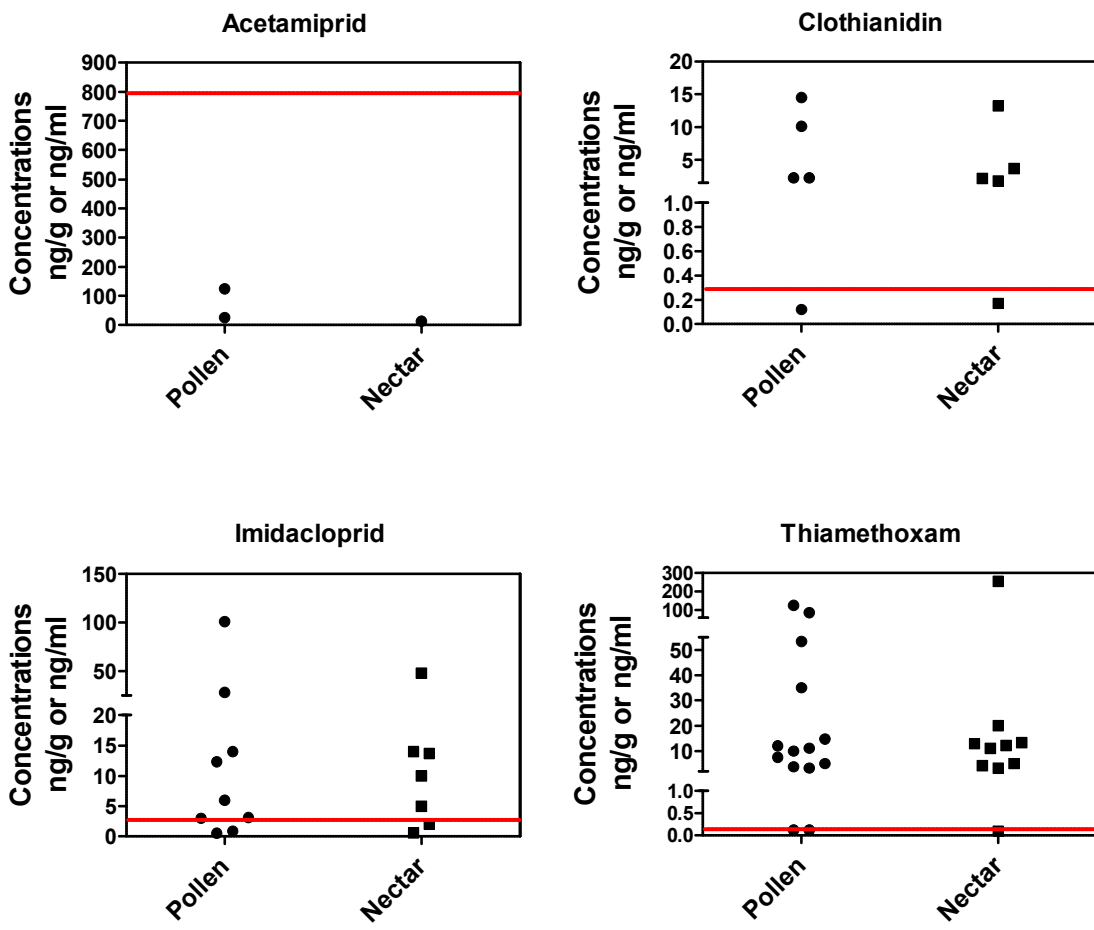
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875 Fig. 6:
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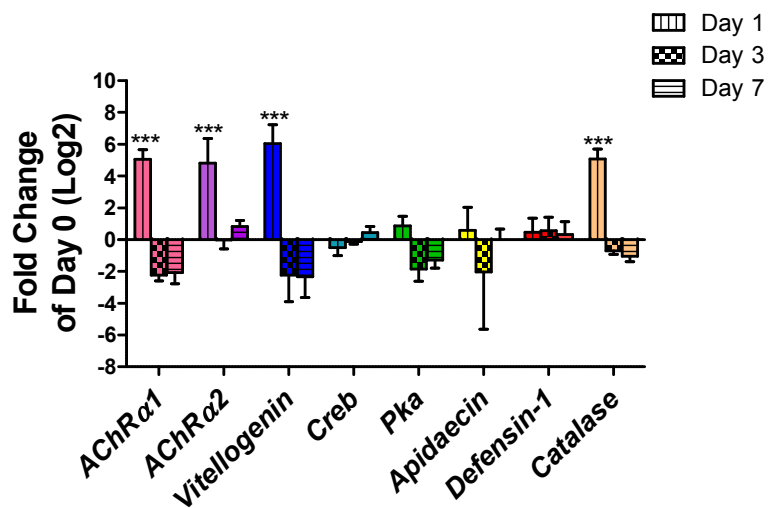
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884 Fig. 7:

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A



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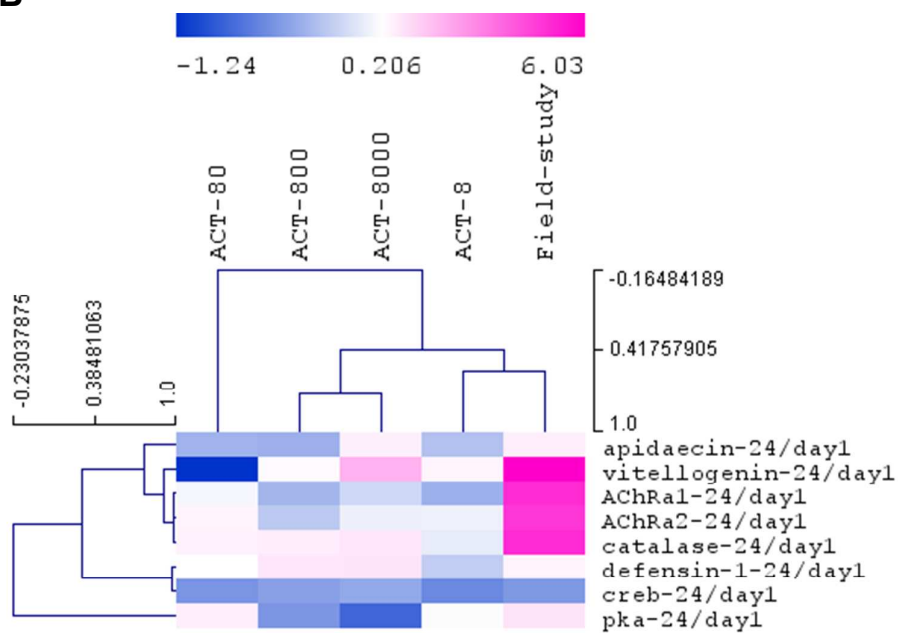
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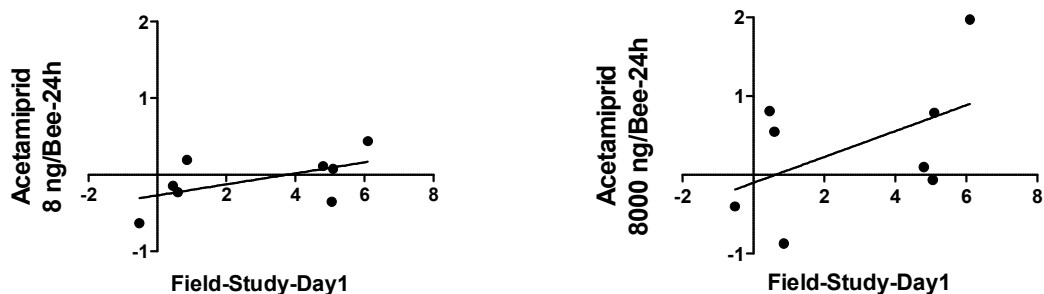
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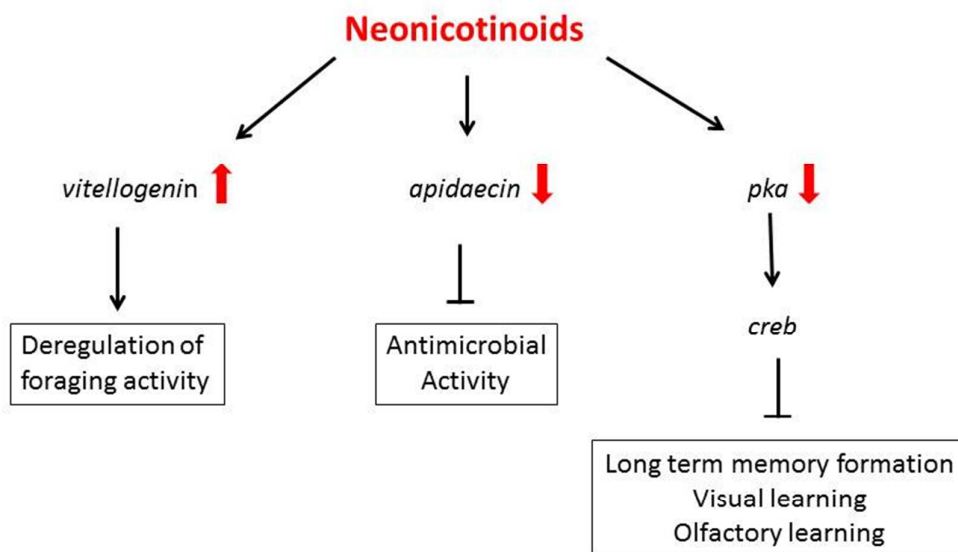
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901 **C**



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908 Fig. 8:
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